

HYDROTHERMAL SYNTHESIS AND ENHANCED PHOTOCATALYTIC ACTIVITY OF $\text{TiO}_2\text{-Fe@CNTs}$ NANOCOMPOSITE FOR METHYLENE BLUE DEGRADATION UNDER VISIBLE LIGHT IRRADIATION

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Abstract. TiO_2 is one of the most attractive metal oxides because of the excellent chemical and photocatalytic properties. However, a problem in the application of TiO_2 is the large band gap energy of 3.2 eV, corresponding to its photocatalytic activity under UV-light irradiation of wavelengths < 387 nm. In this work, TiO_2 nanoparticles doped with iron were grown on the surface of functionalized carbon nanotubes ($\text{TiO}_2\text{-Fe@CNTs}$) to expand the photoabsorbance of the nanocomposite materials in the visible light region and improve their photocatalytic activity. $\text{TiO}_2\text{-Fe@CNTs}$ nanocomposite materials were synthesized by hydrothermal route in Teflon-sealed autoclave at 180°C for 10h. The FE-SEM and X-Ray diffraction measurements were taken for morphology and structural analysis of TiO_2 nanoparticles doped with Fe coating on CNTs. The effects of the iron and CNTs on the enhanced photocatalytic activity for methylene blue degradation under AM 1.5 illumination of 100 mW.cm^{-2} were investigated.

Keywords: TiO_2 , photocatalysis, nanotechnology.

I. INTRODUCTION

Titanium dioxide (TiO_2) has been extensively used semiconductor in solar cells, gas sensors, photocatalysis and photochemical processes due to its stability, low cost and non-toxicity [1]. However, narrow light response range under UV-light irradiation of wavelengths < 387 nm (corresponding to TiO_2 band gap energy about 3.2 eV) which accounts only for 4% of the incoming solar energy and low separation probability of the photoinduced electron – hole pairs in TiO_2 limits its technological applications. Therefore, doping of TiO_2 semiconductor nano-particles with transition metal ions including Fe, Cu, Ni or nonmetal atoms such as N, S, C to shift the optical absorption of from UV to the visible light region has been increasingly studied [2,3]. Doping TiO_2 with other elements can lead to a lower band gap and therefore improve the sunlight utilization of TiO_2 . Additionally, the recombination of electron–hole pairs is suppressed by the introduction of other elements. Recently, the effect of co-doping TiO_2 with both cation and anion on the photocatalytic activity, such as Fe and Eu [4], Zn and Fe [5], N and Ni [6], N and S [7], shows apparently higher photocatalytic activity than that of a single doped TiO_2 . Among the metals doped TiO_2 , Fe doped TiO_2 show greatly improved photocatalytic activity compared to crude TiO_2 . Fe is an

abundant element, environmentally friendly and suitably inserted into the TiO₂ lattice structures, because the ionic radius of iron (0.65 Å) and titanium (0.605 Å) is similar [8]. In this work, we demonstrated a simple hydrothermal method to prepare TiO₂ nanoparticles doped with iron coating on the surface of functionalized carbon nanotubes (TiO₂-Fe@CNTs) nanocomposite material. This approach was expected to improve the photoactivity of TiO₂ in the visible light range by increasing the surface area, creating defect structures to induce space-charge separation and enlarge the absorption region. The obtained materials are characterized by different methods such as: X-ray diffraction (XRD), Field emission scanning electron microscopy (FE-SEM), transmission electron microscopy (TEM), UV-vis spectroscopy. The effects of the Fe and CNTs components on the properties of the TiO₂ were discussed. The photocatalytic activity of the prepared TiO₂-Fe@CNTs composite photocatalysts was evaluated by the degradation of methylene blue under simulated solar light irradiation.

II. EXPERIMENTAL

II.1. Preparation of nanostructured TiO₂ doped with iron

Nanostructured TiO₂ doped with iron (TiO₂-Fe) powder were prepared by sol-gel method, using titanium tetra isopropoxide (Ti[OC₃H₇]₄), Fe(NO₃)₃·9H₂O (dopant), ethanol, HNO₃ and de-ionized water. All the chemicals are of highest purity and commercially available. They were used without further purification. Desired amount of Fe(NO₃)₃·9H₂O (the molar Fe/Ti ratio in the starting materials is 0.6 at.%) was dissolved in the mixture of 17 ml ethanol, 1.6 ml de-ionized water and 0.4 ml concentrated HNO₃. The mixed solution was then slowly dropped into the mixture of 34 ml ethanol and 6 ml titanium tetra isopropoxide, and the final solution was stirred for 30 minutes. The obtained sol was maintained at room temperature for 48 h. The resulting sol was evaporated at 100 °C for 24 hours to gradually form organic brown-colored gel. The yellow or pale white TiO₂-Fe powders were obtained by calcination at 400 °C for 3 hours. Finally, the materials were obtained by filtrating and washing with ethanol and de-ionized water for several times, and then dried at 80 °C for 12 hours.

II.2. Preparation of TiO₂-Fe@CNTs nanocomposite material

TiO₂-Fe@CNTs composites were prepared by a hydrothermal method as follow. Firstly, the CNTs need to be treated with strong acid mixture of 1:3 HNO₃/H₂SO₄ (by volume) under reflux for 12 hours to functionalize surface of CNTs. Due to the inert property of CNTs, this step is necessary to add carboxylic groups (COOH) on to the surface of the CNTs in order to enable them to take part in reactions with the nanoparticles. The resulting solution was filtered and washed thoroughly with water and decantation was used to remove any remaining acid, followed by drying in an oven at 200°C. The obtained powder was designated as acid-functionalized multiwalled carbon nanotubes (CNT-COOH). In the next step, 10 mg of functionalized CNTs was dissolved in a solution of 40ml de-ionized water and 20ml ethanol mixture using ultrasonic treatment for 2h. Then, 6ml titania precursor, namely titanium tetra isopropoxide (Ti[OC₃H₇]₄) solution was dissolved in the mixture of 34 ml ethanol, 0.4 ml concentrated HNO₃, 1.6 ml de-ionized water and 48.2 mg Fe(NO₃)₃·9H₂O. The mixed solution was then added to the functionalized CNTs solution and stirred for another 2 hours at 60°C to get a homogeneous suspension. The suspension was then placed in a 200 ml Teflon-sealed autoclave and maintained at 180°C for 10 hours. Finally, the

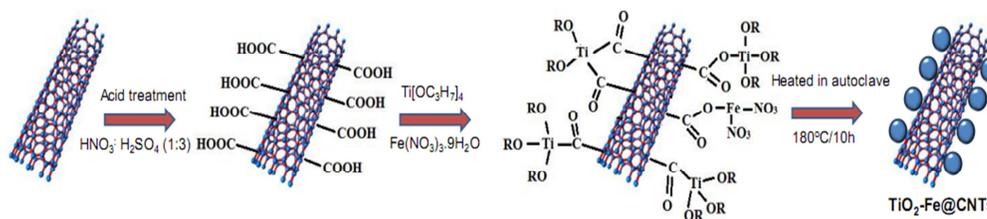


Fig. 1. Synthetic scheme for the preparation of $\text{TiO}_2\text{-Fe@CNTs}$ nanocomposite material

resulting powder was recovered by filtration, rinsed by de-ionized water several times, and dried at 100°C for 12 hours.

II.3. Sample characterization

The surface morphology of samples was investigated by using a “Hitachi S-4800” field emission scanning electron microscopy (FE-SEM) and transmission electron microscopy (TEM, JEOL JEM1010). The crystalline phase was identified by X-ray diffraction (XRD) using a SIEMENS D5000 powder X-ray diffractometer ($\text{CuK}\alpha$ as radiation source, $\lambda=0.15406$ nm). The ultraviolet–visible (UV–vis) absorption spectra were performed by using a Jasco V-670 UV-VIS spectrometer.

II.4. Evaluation of photocatalytic performance

The photoreactor system used in the present work was measured under 1 sun illumination ($100\text{ mW}\cdot\text{cm}^{-2}$, AM 1.5). For a typical experiment, 10 mg of catalyst was put into 20 ml of 0.01 g/l methylene blue (MB) aqueous solution with stirring in darkness for 30 min to reach the adsorption–desorption equilibrium of MB on the catalyst surface prior to illumination. The suspension was then exposed to light with continuous magnetic stirring. Then suspensions were collected and centrifugally separated every 20 min, and the filtrate was returned to the vial for measurement of the concentration MB degradation using a UV–vis spectrophotometer at a wavelength of 664 nm.

III. RESULTS AND DISCUSSION

III.1. Structure and morphology

Morphological characterization of $\text{TiO}_2\text{-Fe@CNTs}$ nanocomposite material was carried out by FE-SEM (Fig. 2a) and TEM (Fig. 2b). As shown in Fig. 2a, FE-SEM image of $\text{TiO}_2\text{-Fe@CNTs}$ nanocomposite material exhibits the presence of nanoparticles on the surface of carbon nanotubes. Some clusters of 100 nm nanoparticles were either randomly immobilized on the surface of CNTs or densely collected among CNTs bundles. TEM image in Fig. 2b reveals that granular $\text{TiO}_2\text{-Fe}$ nanoparticles with 7 nm in diameter was intimately bound to the surface of CNTs.

Fig. 3 shows the XRD characterization of the $\text{TiO}_2\text{-Fe}$ nanoparticles and $\text{TiO}_2\text{-Fe@CNTs}$ nanocomposite material. The XRD pattern also shows that intensities of 0.6 at.% Fe doped TiO_2 (Fig. 3a) were slightly higher than $\text{TiO}_2\text{-Fe@CNTs}$ (Fig. 3b). As seen from Fig. 3b, the strong diffraction peak for the $\text{TiO}_2\text{-Fe@CNTs}$ at the angle (2θ) of 25.8° can be assigned to the C (002) reflection of the hexagonal graphite structure. There are intensity peaks corresponding to anatase

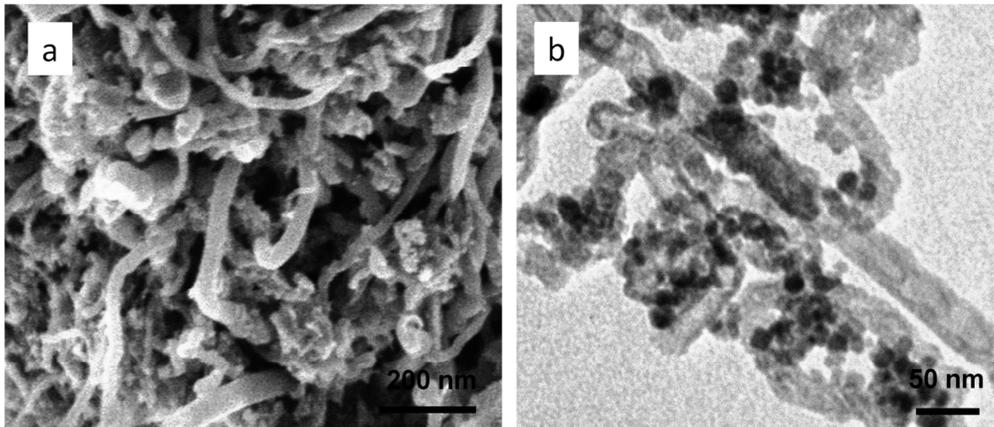


Fig. 2. FE-SEM (a) and TEM (b) images of the TiO₂-Fe@CNTs nanocomposite material

phase of TiO₂ for all samples without the appearance of trace rutile phase. No Fe peak was observed for either Fe doped TiO₂ or TiO₂-Fe@CNTs composite. This indicates that doping Fe ions with small percentages did not affect the crystalline structure of TiO₂ nanoparticles and no precipitate formation of doped metal such as iron oxides or Fe_xTiO_y. As noted previously [8,9], Fe³⁺ can be suitably inserted into the TiO₂ lattice structures and substituted Ti⁴⁺ site because the ionic radius of Fe³⁺ and Ti⁴⁺ is similar.

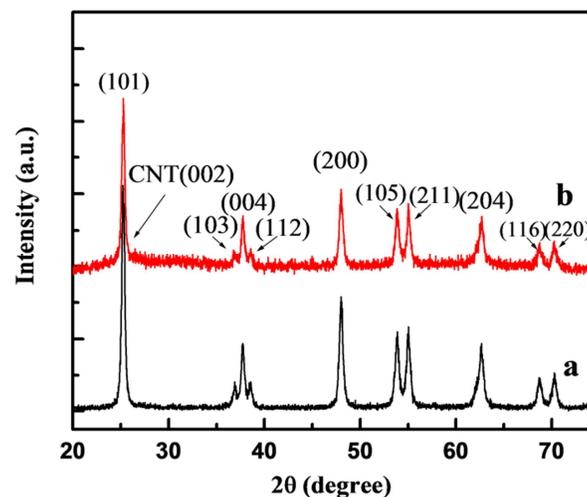


Fig. 3. XRD patterns of the TiO₂-Fe nanoparticles(a) and TiO₂-Fe@CNTs nanocomposite material (b).

III.2. Optical properties

To study the optical properties of Fe doped TiO₂ and TiO₂-Fe@CNTs samples, the UV–vis absorption spectra were measured, as shown in Fig. 4a. Compared with pure anatase TiO₂ that can only absorb $\lambda \leq 387$ nm UV light, the light absorption edges of Fe doped TiO₂ and TiO₂-Fe@CNTs nanocomposites move remarkably with a red shift to the visible range, as well as reduction in the bandgap energy.

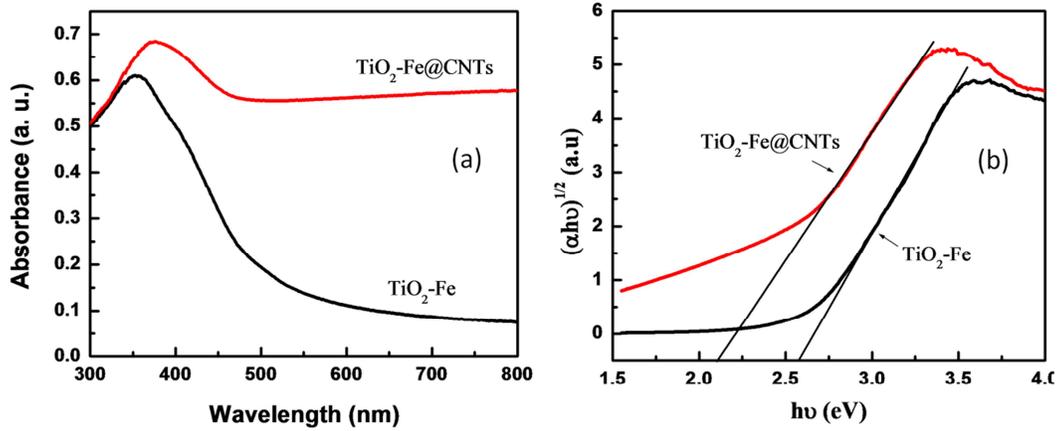


Fig. 4. UV–visible spectra (a) and plots of $(\alpha h\nu)^{1/2}$ as a function of photon energy (b) of the TiO₂-Fe nanoparticles and TiO₂-Fe@CNTs nanocomposite material.

The band gap energy E_g and absorption coefficient α are related by the following equation [10]:

$$\alpha h\nu = A(h\nu - E_g)^m$$

Where α is the absorption coefficient, $h\nu$ is the photon energy in eV, and E_g is the band gap energy in eV. A is a constant related to the effective mass of the electrons and holes and m being equal to 0.5 for allowed direct transition and 2 for an allowed indirect transition. Plots of between $(\alpha h\nu)^{1/2}$ and photon energy ($h\nu$) for the TiO₂-Fe nanoparticles and TiO₂-Fe@CNTs nanocomposite material shown in the Fig. 4b. The energy gap E_g of the samples was evaluated from the intercept of the linear portion of the each curve with the $h\nu$ in X-axis. The straight lines imply that the TiO₂-Fe and TiO₂-Fe@CNTs nanocomposite samples have energy band gaps of 2.6 eV and 2.1 eV, respectively. The red shift in absorption spectrum and reduction in bandgap are attributed to the transfer of 3d-electrons from Fe to the conduction band of TiO₂.

III.3. Photocatalytic degradation of methylene blue

The photocatalytic degradation efficiency ($\eta\%$) could be calculated as the following equation:

$$\eta(\%) = \frac{C_o - C_t}{C_o} \times 100$$

where C_o is the initial absorbance of the MB solution and C_t , its absorbance at different irradiation times. All C values were obtained by the maximum absorption at 664 nm in the absorption spectrum in order to evaluate the degradation efficiency as shown in Fig. 5.

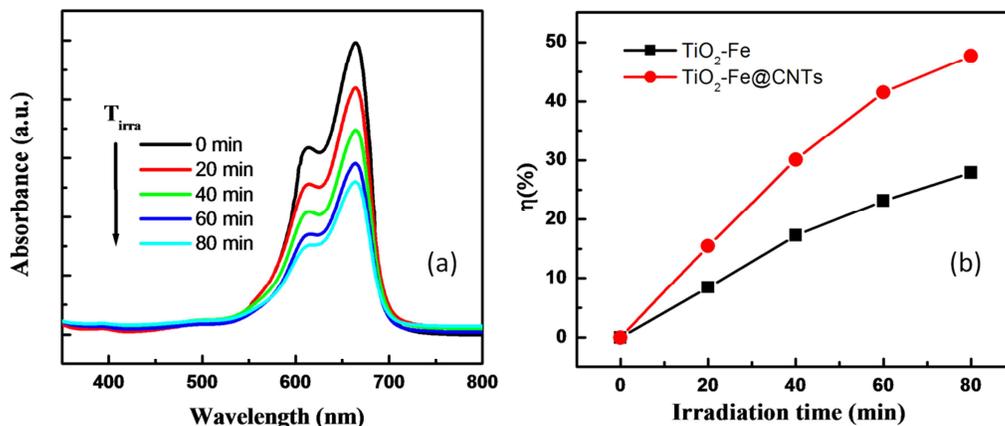


Fig. 5. Degradation indicated by absorption spectrum against TiO₂-Fe@CNTs photocatalyst (a) and the percentage degradation of methylene blue solutions under simulated solar light irradiation (b).

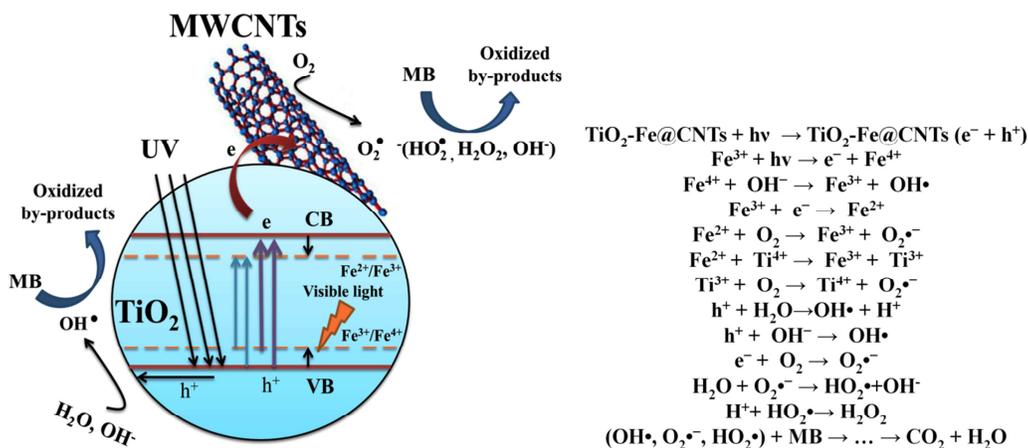


Fig. 6. Schematic diagram of the separation of generated electrons and holes on the interface of TiO₂-Fe@CNTs photocatalyst under simulated solar light irradiation.

Enhanced activity of the TiO₂-Fe@CNTs photocatalyst in the visible light region as seen in the Fig. 5 is attributed to the formation of appearance of localized electronic gap states ascribed to Fe d-d or “localized charge transfer” transitions simultaneously. In addition, Ti-C and Ti-O-C defect sites in the TiO₂-Fe@CNTs photocatalyst act as trapping sites for photogenerated charges. When the catalyst is irradiated by photons, electrons (e⁻) are excited from the valence band (VB) to the conduction band (CB) of Fe doped TiO₂ nanoparticles creating holes (h⁺) in the VB. The hole (h⁺) on the metal oxide nanoparticles oxidize hydroxyl groups to form hydroxyl radical (OH[·]) which can decompose the methylene blue (MB). While CNTs act as a photo-generated electron

acceptor to promote interfacial electron transfer process and reduce the electron-hole recombination [11]. The adsorbed oxygen molecules on the surface of CNTs react with the electrons forming very reactive superoxide radical ion (O_2^-) which oxidize methylene blue (MB). The proposed photocatalytic methylene blue degradation mechanisms were described in the Fig. 6 and the reactions could be expressed as follows:

IV. CONCLUSIONS

TiO₂-Fe@CNTs nanocomposite photocatalyst was successfully prepared via hydrothermal synthesis. The structure, morphology and optical properties were investigated. The TiO₂-Fe@CNTs nanocomposite photocatalysts showed a high efficiency in the degradation of methylene blue under visible light. A tight interface between CNTs and TiO₂ doped with Fe was created. Doping TiO₂ with Fe ions can also expand its light absorption range to visible light through the formation of mid-gap states. On the other hand, CNTs with good adsorption capacity was successfully bonded with TiO₂ to reduce the electron-hole recombination rate and prevent the particles agglomeration. The higher visible light photocatalytic activity for methylene blue degradation of the TiO₂-Fe@CNTs nanocomposite was attributed to the increased adsorption capacity, enhanced light absorption and better charge separation. This semiconductor heterogeneous photocatalysis based on the advanced oxidation process (AOP) has enormous potential for the treatment of a wide range of organic contaminants in water.

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